

1 **Impact of distillers' grains moisture and inclusion level on greenhouse gas emissions in the**  
2 **corn-ethanol-livestock life cycle**

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18 ABSTRACT

19 A model was previously developed (Biofuel Energy Systems Simulator; BESS;  
20 [www.bess.unl.edu](http://www.bess.unl.edu)) to predict greenhouse gas (GHG) emissions and net energy yield when  
21 ethanol is produced from corn. The model predicts animal performance and subsequently the  
22 feed replacement value of the ethanol co-products. The BESS model was used to develop  
23 several scenarios over a range of livestock classes and amount of drying of co-products. New  
24 meta analysis equations for performance of feedlot cattle fed 0 to 500 g/kg of dietary dry matter  
25 (DM) as corn wet (WDGS, 320 g/kg DM), modified (MDGS, partially dried WDGS, 460 g/kg  
26 DM), or dry (DDGS, 900 g/kg DM) distillers' grains plus solubles (DGS) replacing dry rolled  
27 (DRC) and high moisture corn (HMC) were incorporated into the model. Equations were  
28 derived from pen-level performance for 20 trials evaluating WDGS, 4 trials evaluating MDGS,  
29 and 4 trials evaluating DDGS conducted at University of Nebraska research feedlots. Feeding  
30 value of WDGS was 145 to 131% of the corn replaced in diets at levels from 200 to 400 g/kg of  
31 diet DM. Using the same approach, feeding value of MDGS was 124 to 117% and 110 to 112%  
32 for DDGS. Midwest corn-ethanol-livestock life cycle GHG reduction relative to gasoline (97.7 g  
33 CO<sub>2</sub>e/MJ ethanol) was greatest when WDGS was fed to feedlot cattle and decreased from 61 to  
34 57% for 200 to 400 g/kg of diet DM as WDGS. Feeding MDGS and DDGS to feedlot cattle  
35 resulted in a reduction of GHG emissions by 53 to 50% and 46 to 41%, respectively. Life cycle  
36 GHG reduction for WDGS, MDGS, or DDGS for dairy cows was 53, 48, and 43%, respectively,  
37 and DDGS for swine was 42%. Reduction in GHG emissions are less when DDGS are fed  
38 compared to WDGS and MDGS for beef or dairy. Reduction in GHG emissions was comparable  
39 for all three livestock classes when DDGS was fed. Partial drying (MDGS) or complete drying  
40 (DDGS) of WDGS reduced both feeding value and GHG reductions for corn-ethanol relative to  
41 gasoline. Feeding WDGS to feedlot cattle was the optimum feed use of DGS based on feeding

42 performance and GHG reduction. Accurate data for ethanol and gasoline GHG emissions are  
43 essential for providing a meaningful comparison of these fuels.

44 *Keywords:* Distillers' Grains, cattle performance, greenhouse gases, life cycle assessment

45 *Abbreviations:* ADG, average daily gain; BESS, Biofuel Energy Systems Simulator; DGS,  
46 distillers' grains plus solubles; DDGS, dried distillers' grains plus solubles; DM, dry matter;  
47 DMI, dry matter intake; gCO<sub>2</sub>e, grams of carbon dioxide equivalents; G:F, feed efficiency;  
48 GHG, greenhouse gas; MDGS, modified distillers' grains plus solubles; WDGS, wet distillers'  
49 grains plus soluble.

## 50 **1. Introduction**

51 Corn (*Zea mays*) distillers' grains plus solubles (DGS) are an important part of the corn-  
52 ethanol-livestock life cycle when comparing greenhouse gas (GHG) emissions of ethanol to  
53 gasoline. Distillers' grains contain a significant quantity of energy and offsets corn, urea and  
54 soybean meal in livestock diets. The energy and protein replacement value of DGS is  
55 dependent on dietary inclusion level and livestock class fed. Ethanol plant energy use and  
56 associated GHG emissions are affected by moisture content of DGS produced. All ethanol  
57 plants produce wet DGS (WDGS; 650-690 g/kg moisture). Some plants choose to remove  
58 moisture from WDGS to form modified DGS (MDGS; 520-580 g/kg moisture) or dried DGS  
59 (DDGS; 80-120 g/kg moisture). Ethanol plant energy use (*e.g.*, natural gas) to remove moisture  
60 has been identified as a parameter of importance in comparing GHG emissions from ethanol  
61 relative to gasoline (Liska et al., 2009; Bremer et al., 2010b).

62           The Biofuel Energy Systems Simulator (BESS; [www.bess.unl.edu](http://www.bess.unl.edu)) was developed to  
63 compare life cycle GHG emissions from ethanol production relative to gasoline as a motor fuel,  
64 while accounting for the dynamic interactions of corn production, ethanol plant operation, and  
65 co-product feeding to livestock. The co-products fed to livestock replace other feedstuffs that  
66 supply protein, energy and minerals. The replacement values vary with livestock class and  
67 nutrient, such as protein, that is replaced. Modeling GHG emissions requires accurate biological  
68 equations derived from animal performance over a broad range of DGS feeding conditions.  
69 Good summaries of DGS feeding to swine and dairy cattle were available, but limited data on  
70 feeding DGS to feedlot cattle were available. The BESS model was developed with feedlot  
71 cattle performance equations available at the time. A meta-analysis using a limited number of  
72 experiments was used to develop equations for feeding of WDGS while individual feeding trials  
73 of MDGS and DDGS were used (Liska et al., 2009; Bremer et al., 2010b). Multiple trials for all  
74 three DGS moistures have been completed more recently to augment the initial datasets.  
75 Revised meta-analyses of cattle performance equations developed from these more complete  
76 databases should improve the accuracy of modeling GHG emissions from ethanol production.

77           Therefore, the objectives of the current study were to update equations predicting cattle  
78 performance when fed WDGS, MDGS, and DDGS in BESS with the most complete data  
79 available and to evaluate the impact of DGS moisture and inclusion level in livestock diets on  
80 ethanol GHG emissions from the corn-ethanol-livestock life cycle relative to gasoline.

## 81 **2. Materials and methods**

### 82 *2.1 Cattle performance data*

83 Several scenarios were developed using the BESS model in order to compare GHG  
84 reduction when ethanol is produced compared to gasoline. The scenarios were developed over a  
85 range of livestock classes and amount of drying of co-products. Feedlot cattle were evaluated  
86 when fed levels (100-400 g/kg) of WDGS, MDGS, or DDGS. Lactating dairy cows were  
87 evaluated when fed WDGS, MDGS, and DDGS. Swine were evaluated only when fed DDGS.

88 *2.2 Dataset* Prior to conducting the analysis using the BESS model to estimate GHG  
89 emissions in the various scenarios, a new dataset was developed to update the equations in the  
90 BESS model. Prediction equations of performance of cattle fed wet DGS were developed from  
91 20 feedlot cattle finishing trials with 350 pen means, representing 3,365 steers fed (Larson et al.,  
92 1993; Ham et al., 1994; Al-Suwaiegh et al., 2002; Vander Pol et al., 2005; Godsey et al., 2008a,  
93 2008b; Meyer et al., 2008; Wilken et al., 2008; Corrigan et al., 2009; Rich et al., 2009; Vander  
94 Pol et al., 2009; Loza et al., 2010; Luebbe et al., 2010; Moore et al., 2010; Nuttelman et al.,  
95 2010; Rich et al., 2010; Sarturi et al., 2010). Equations for modified DGS were also developed  
96 from four University of Nebraska-Lincoln (UNL) feedlot trials with 85 pens, representing 680  
97 steers (Adams et al., 2007; Huls et al., 2008; Luebbe et al., 2010; Nuttelman et al., 2010).  
98 Equations to predict performance of cattle fed wet DGS were developed from four UNL feedlot  
99 trials with 66 pens, representing 581 steers (Ham et al., 1994; Buckner et al., 2010; Nuttelman et  
100 al., 2010; Sarturi et al., 2010). In all trials cattle performance traits measured included average  
101 daily gain (ADG), dry matter intake (DMI), feed efficiency (G:F), hot carcass weight (HCW),  
102 ribeye area (REA), and 12<sup>th</sup> rib fat thickness.

103 All trials included in the analyses evaluated feeding corn DGS replacing dry-rolled corn,  
104 high-moisture corn, or a blend of the two corn types. Individual animal carcass data were  
105 collected on all steers and feeding performance was calculated from carcass adjusted final

106 weight. In each trial a single DGS moisture type co-product was fed as 0 to 500 g/kg DM in the  
107 diet. All trials were conducted under similarly managed feedlot research settings at University  
108 of Nebraska Beef Research Feedlots. Animal use procedures were reviewed and approved by  
109 the University of Nebraska Institutional Animal Care and Use Committee.

### 110 *2.3 Data analysis*

111 In order to develop equations to update the BESS model, it was necessary to combine  
112 data from the various experiments. Meta-analysis methodology for integrating quantitative  
113 findings from multiple studies was utilized for data analysis of the three DGS products (St-  
114 Pierre, 2001). This method accounts for the random effect of individual trials with a structured  
115 iterative analytical process using the PROC MIXED procedure of SAS (SAS Inst., Inc., Cary,  
116 NC, USA). Pen mean was the experimental unit of analysis. Trials were weighted by number of  
117 WDGS levels to prevent artificial linear responses from trials with 0 and one other level of DGS  
118 evaluated. Each DGS moisture type was analyzed with a separate dataset. Biological  
119 performance equations were developed based on significant model variables. Variables tested  
120 included ADG, DMI, and G:F. The intercepts (no DGS diet) of the MDGS and DDGS predicted  
121 performance equations were scaled to the intercept of the WDGS dataset to compare differences  
122 in cattle performance relative to a common no DGS diet. A common no DGS diet in the  
123 Midwestern United States would include a combination of DRC and HMC at 800-900 g/kg of  
124 the diet DM, DGS is added to the diet to replace corn. The equation adjustment allowed the  
125 evaluation of how an individual steer would perform if given one of the three products relative to  
126 a common base point.

### 127 *2.4 Model parameters*

128           The assumptions and calculations of BESS have been discussed extensively (Liska et al.,  
129 2009, Bremer et al., 2010b). The BESS model estimates the net energy yield from producing  
130 ethanol (ethanol plus co-product credits minus energy inputs). The model also estimates GHG  
131 emissions (CO<sub>2</sub> equivalents) from non-renewable energy sources needed to produce the corn,  
132 produce the ethanol, dry co-products (if needed) and transport co-products to feedlots. The GHG  
133 emissions are then compared to that from gasoline used as liquid fuel and reported as percentage  
134 reduction. Energy costs and GHG emissions with production of fossil fuels, fertilizer inputs and  
135 electricity are included. Non fossil fuel GHG emissions include N<sub>2</sub>O from fertilizer and manure.  
136 Bremer et al. (2010b) further discussed the dynamic livestock and DGS components of the BESS  
137 model. Midwestern United States corn production efficiency of 362 g CO<sub>2</sub>eq/kg of corn DM  
138 was used for all scenarios (Bremer et al., 2010b). Ethanol plant GHG emissions from ethanol  
139 production and dryer operation were developed from a survey of 9 ethanol plants (Bremer et al.,  
140 2010b). Average ethanol plant GHG emissions from natural gas and electricity use for plant  
141 operation and DGS drying were 21.0, 25.6, and 30.5 g CO<sub>2</sub>e/MJ ethanol for WDGS, MDGS, and  
142 DDGS, respectively.

143           We have assumed that livestock production would be similar in quantity whether DGS is  
144 fed or if not. Therefore, a partial budget approach was used assuming GHG emissions would be  
145 the same except for any direct effects the use of DGS would have on GHG emissions, such as  
146 transportation. It was further assumed that CO<sub>2</sub> originating from corn and converted to CO<sub>2</sub>,  
147 either in the biorefinery or by the livestock, is resynthesized into organic, energy containing  
148 nutrients the subsequent growing season (renewable energy). McGinn et al. (2009) reported that  
149 feeding DDGS in forage based growing diets reduced methane emissions. This was also shown  
150 by Behlke et al. (2007) in lambs. However, Behlke et al. (2007) showed greater methane

151 emissions in lambs fed high grain diets when DGS was included, even though the unsaturated  
152 lipid in the DGS would be expected to reduce methane emissions. Alternatively, the digestion of  
153 the fiber in DGS compared to starch in the corn replaced would be expected to increase methane  
154 production. Because of these conflicting reports and mechanisms, we chose to assume similar  
155 methane production by cattle fed diets containing DGS and those containing primarily corn.

156 The BESS model contains equations that predict animal DMI and G:F. These equations  
157 are used to predict cattle growth to a common weight (Bremer et al., 2010b). The equations for  
158 WDGS, MDGS, and DDGS were updated based on the meta-analyses previously described.  
159 Distillers' grains replaced corn and urea nitrogen in beef finishing diets (Klopfenstein et al.,  
160 2008a), but replaced corn and soybean meal in swine finishing and dairy lactating diets (Bremer  
161 et al., 2010b). Summaries of dairy and swine DGS feeding data (Schingoethe, 2008; Stein,  
162 2008) do not indicate a feeding value of DGS greater than a combination of soybean meal and  
163 corn. Therefore, a direct replacement of corn and soybean meal (kg for kg of DM) was utilized  
164 when DGS is fed to these animal classes.

165 An average emissions intensity for gasoline considering a tar sands fraction (7/100  
166 barrels of petroleum) and California reformulated gasoline blendstock is estimated at 97.7  
167 gCO<sub>2</sub>e/MJ. This value was used as the gasoline reference point for all scenarios (Liska and  
168 Perrin, 2009).

### 169 *2.5 Scenarios evaluated*

170 Corn production efficiency and ethanol plant operation except for drying of DGS was  
171 held constant for all scenarios. Greenhouse gas emissions of ethanol produced from the corn-  
172 ethanol-livestock life cycle relative to gasoline were calculated for the following scenarios. The

173 ethanol plant produces WDGS fed at 100, 200, 300, or 400 g/kg of diet DM to feedlot cattle or  
174 fed at 100, 200, or 300 g/kg of diet DM to lactating dairy cows. Similar scenarios for both  
175 feedlot and dairy were evaluated for MDGS and DDGS. Swine use of DGS is limited to DDGS  
176 and scenarios of 90,180, or 270 g/kg of finishing diet DM were evaluated.

### 177 **3. Results**

#### 178 *3.1 Dataset*

179 Steer DMI increased quadratically as DGS inclusion level increased (Table 1). The  
180 greatest improvement in DMI occurred when DDGS replaced corn (DRC and HMC). The DMI  
181 response to MDGS inclusion was intermediate to DDGS and WDGS. Maximum DMI of steers  
182 fed DDGS occurred at a greater level of DGS inclusion than MDGS, and the maximum DMI of  
183 steers fed WDGS occurred at the lowest level of DGS inclusion of the three DGS moisture  
184 products. Quadratic increases in ADG and G:F were observed when steers were fed WDGS or  
185 MDGS. Steer ADG and G:F improved linearly as DDGS replaced corn in the diet. Steer ADG  
186 was similar for the three DGS moisture products. All DGS products contained greater feeding  
187 value than corn, measured by the increase in feed efficiency of DGS diets compared to corn  
188 based diets. The feeding values of WDGS, MDGS, and DDGS, when fed at 200 to 400 g/kg of  
189 diet DM, were 143 to 130, 124 to 117, and a constant 112% of corn (DM basis), respectively.  
190 The feeding value of DGS decreased as moisture level decreases. The feeding value of WDGS  
191 and MDGS decreased as inclusion level increases. The feeding value of DDGS was a constant  
192 112% of corn DM.

#### 193 *3.2 GHG emissions of ethanol*

194 All scenarios evaluated had ethanol life cycle emissions less than gasoline (Table 2). The  
195 life cycle that included feeding WDGS to feedlot cattle had the least ethanol GHG emissions of  
196 the scenarios evaluated. The next best option was feeding WDGS to dairy cows. Feeding MDGS  
197 to feedlot cattle created fewer GHG emissions than feeding MDGS or DDGS to dairy cattle.  
198 Feeding DDGS to feedlot cattle had slightly fewer GHG emissions than feeding DDGS to swine  
199 and dairy cows.

## 200 **4. Discussion**

201

### 202 *4.1 DGS moisture*

203 A decrease in steer performance as moisture is removed from WDGS, as indicated by the  
204 results of the meta-analyses, is in agreement with individual studies that evaluated both WDGS  
205 and DDGS (Ham et al., 1994; Sarturi et al., 2010; Nuttelman et al., 2010). Those three studies  
206 evaluated feeding DGS in the WDGS or DDGS forms and found the feeding value of WDGS to  
207 be greater than that of DDGS. Nuttelman et al. (2010) conducted the first study to evaluate  
208 feeding multiple dietary inclusion levels of WDGS, MDGS, and DDGS in the same trial. In  
209 addition, the MDGS and DDGS were sourced from the same ethanol plant and all 3 types of  
210 DGS were identical in nutrient composition. The researchers observed the feeding value of  
211 WDGS to be greater than MDGS, which were both greater than DDGS. Their results indicate  
212 cattle fed drier DGS products eat to a constant energy intake, which are supported by the  
213 findings from our study that DMI of steers increased as DGS moisture decreased, when  
214 compared at the same ADG.

215           The feeding value of DGS is set at the ethanol plant with management decisions on how  
216 to market WDGS. Target market livestock populations and DGS transportation costs are drivers  
217 of how WDGS is processed at the ethanol plant (Buckner et al., 2008; Bremer et al., 2010b).  
218 Drying WDGS improves shelf life and decreases shipping costs due to less moisture being  
219 hauled. Drying DGS allows access to markets unattainable with WDGS by moving DDGS into  
220 export markets, the swine industry, and livestock industries in other regions within North  
221 America. This flexibility comes at a cost. In addition to the decrease in feeding value of DDGS  
222 relative to WDGS, the fixed and variable cost of owning and operating a dryer in an ethanol  
223 plant are significant (Baumel, 2008). Ethanol plant decisions on DGS moisture management  
224 also impact the GHG balance of ethanol produced. Ethanol plants producing DDGS require  
225 167% of the energy and produce 145% of the GHG emissions of ethanol plants producing  
226 WDGS (Liska et al., 2009). This emphasizes making ethanol production decisions that are  
227 economically and environmentally sound.

228           Clearly, the greatest reduction of GHG emissions occurred when WDGS were fed to  
229 feedlot cattle (beef). This is because of the greater replacement value of the co-product for beef  
230 compared to other livestock types and because drying is not required. Dairy cattle have slightly  
231 lower GHG reductions than feedlot cattle because there is less replacement value of the co-  
232 products. All livestock types had lower reduction in GHG emissions when DDGS was fed,  
233 obviously because of the use of fossil fuel to dry the co-product.

#### 234 *4.2 Gasoline reference point*

235           The evaluation of ethanol relative to gasoline not only requires accurate evaluation of the  
236 ethanol production cycle, but also an accurate reference point for the GHG-intensity of gasoline.

237 Gasoline emissions not only include combustion emissions, but also upstream emissions from  
238 crude oil recovery, refinery emission, and flaring losses (Brandt and Farrell, 2007). Emissions  
239 due to military security associated with acquisition of Middle Eastern petroleum, changes in the  
240 composition of petroleum supplies toward more GHG-intensive fuels, and other additional  
241 emissions from petroleum processing must also be considered (Liska and Perrin, 2009). Indirect  
242 GHG emissions from military security for maritime oil transit are estimated to raise the GHG  
243 intensity of gasoline from the Middle East by roughly 20% over the conventional baseline (Liska  
244 and Perrin 2010).

245 Ethanol production does not displace average gasoline, but displaces a marginal unit of  
246 gasoline that may have a much greater environmental cost than average gasoline (US EPA,  
247 2010). As the proportion of gasoline derived from more energy intense processes increases, the  
248 GHG life cycle reference point of gasoline should be updated to compare a marginal liter of  
249 gasoline to an equal energy quantity from ethanol. The GHG-intensity of gasoline is increasing  
250 due to depletion of efficiently accessible deposits (Brandt and Farrell, 2007). Unconventional  
251 and less efficiently processed sources of petroleum such as tar sands, coal-to-liquids, and oil  
252 shale will likely be used to fill the difference between current petroleum supply and energy  
253 demand. In fact, Canadian tar sands could supply 20 out of every 100 barrels of US gasoline by  
254 2020 (Liska and Perrin, 2009).

#### 255 *4.3 Indirect GHG impacts of ethanol and gasoline*

256 Indirect GHG emissions from ethanol and gasoline, such as land use change, were not  
257 evaluated in this study due to the immense complexity in calculating the totality of significant  
258 indirect GHG emissions (Liska and Perrin, 2009; US EPA, 2010). A methodology to incorporate

259 both reasonably accurate scientific knowledge about direct life cycle emissions and relatively  
260 diffuse and uncertain scientific knowledge concerning potentially significant indirect emissions  
261 must be developed to fully evaluate the GHG mitigation potential of ethanol (Liska and Perrin,  
262 2009; US EPA, 2010). This is especially true when the indirect effects may provide a large  
263 impact on the life cycle being analyzed.

264           One may be tempted to add the single indirect emission from land use change due to  
265 increased ethanol production (Searchinger et al., 2008), yet land use change is only one  
266 significant indirect GHG emission among many. Other significant indirect emissions include  
267 military security emissions, changes in rice cultivation, and changes in livestock globally (Liska  
268 and Perrin, 2009; Liska and Perrin 2010; US EPA, 2010). Further research is needed before we  
269 can have reasonable confidence in the net effects of indirect GHG emissions of both biofuels and  
270 petroleum fuels (Liska and Perrin, 2009). A comprehensive assessment of the total GHG  
271 emissions implications of substituting ethanol for petroleum needs to be completed before the  
272 impact of indirect GHG emissions from land use change alone can be accurately determined.

#### 273 *4.4 Current ethanol production vs. future expansion GHG emissions*

274           Indirect land use change is associated primarily with future expansion of the ethanol  
275 industry. Emissions from existing ethanol production facilities are limited to direct emissions,  
276 given whatever indirect emissions were associated with initiating ethanol production at these  
277 facilities has already occurred. Because of this, biofuels use now from existing facilities not only  
278 reduces GHG emissions from transportation fuel use compared to petroleum, but also supports  
279 national security goals and rural development objectives. Evaluation of these additional policy

280 objectives are not considered in GHG emissions modeling frameworks, but are important  
281 considerations when comparing fuels.

#### 282 4.5 Future co-products

283 Distillers' grains are used not only as a protein source but also as an energy source by  
284 feedlot cattle (Klopfenstein et al., 2008a; NRC, 1996). Ruminants are able to utilize the fat, fiber,  
285 and protein components of DGS. Utilization of protein as a source of energy increases N  
286 excretion and  $\text{NH}_3$  emissions, however. Fractionation of DGS products for biodiesel production  
287 from the fat component and cellulosic ethanol production of the fiber fraction will result in a  
288 concentrated protein source. The GHG balance of ethanol and other co-products produced from  
289 fractionated corn processes may be significantly different from the current systems analyzed due  
290 to uses of co-products produced, change in corn processing, and environmental costs of  
291 implementing the technology. The feeding value of these products may also be reduced (Buckner  
292 et al., 2010). Furthermore, exploitation of fibrous biomass fermentation for ethanol production  
293 would directly compete for the resource niche that cattle currently utilize.

294 Although ethanol production has altered the availability of corn for livestock production,  
295 the use of DGS as livestock feed has helped to maintain the synergistic relationship between the  
296 livestock and corn production industries. Feeding DGS results in up to 0.43 kg of corn DM offset  
297 as DGS for each kg of corn DM fermented at the ethanol plant. The US livestock industry is of  
298 sufficient scope to fully utilize DGS production from a 69 billion liters per year corn ethanol  
299 industry (Bremer et al., 2010b). That is a corn ethanol industry 1.7 times larger than the current  
300 40 billion liters per year ethanol production capacity (RFA, 2009). These DGS use calculations  
301 are conservative since they do not account for exporting DGS and feeding DGS to non-lactating

302 dairy cows, beef cattle on grass, feedlot cattle finished in yards less than 1,000 cattle capacity,  
303 and poultry (Klopfenstein et al., 2008b). Increasing the scope of corn ethanol production would  
304 not significantly alter ethanol GHG emissions of individual plants (Bremer et al., 2010b).

## 305 **5. Conclusion**

306 Feeding DGS to livestock is a significant contribution to the environmental benefit of fuel  
307 ethanol relative to gasoline. The GHG emissions benefits of ethanol are determined by how  
308 DGS moisture is managed at the ethanol plant and what animal classes are fed DGS. Feeding  
309 WDGS to feedlot cattle provided the optimum feed use of DGS for livestock. Partial drying  
310 (MDGS) or complete drying (DDGS) of WDGS reduced the feeding value and increased ethanol  
311 GHG emissions relative to WDGS. In state and federal GHG regulations for fuels, regulators  
312 must continually update and use the most representative and accurate data for assessing ethanol  
313 and gasoline GHG emissions. Yet, achieving this accuracy requires much more complete  
314 research on the underlying systems involved, such as the research results presented here.

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- 432

433 Table 1.

434 Finishing steer performance when fed different dietary inclusions of corn wet distillers' grains plus solubles (WDGS<sup>a</sup>), modified  
 435 distillers' grains plus solubles (MDGS) or dried distillers' grains plus solubles (DDGS) replacing dry rolled and high moisture corn.

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436 DGS Inclusion <sup>b</sup> :	0DGS	10DGS	20DGS	30DGS	40DGS	Lin <sup>c</sup>	Quad <sup>c</sup>
437							
438 WDGS <sup>d</sup>							
439 DMI <sup>a</sup> , kg/d	10.4	10.6	10.6	10.4	10.2	0.01	< 0.01
440 ADG <sup>a</sup> , kg	1.60	1.71	1.77	1.78	1.75	< 0.01	< 0.01
441 G:F <sup>a</sup>	0.155	0.162	0.168	0.171	0.173	< 0.01	< 0.01
442 Feeding value, % <sup>e</sup>		150	143	136	130		
443 MDGS <sup>f</sup>							
444 DMI, kg/d	10.4	10.8	10.9	10.9	10.6	0.95	< 0.01
445 ADG, kg	1.60	1.71	1.77	1.78	1.74	< 0.01	< 0.01

446	G:F	0.155	0.159	0.162	0.164	0.165	< 0.01	0.05
447	Feeding value, % <sup>e</sup>		128	124	120	117		
448	DDGS <sup>f</sup>							
449	DMI, kg/d	10.4	10.9	11.2	11.3	11.3	< 0.01	0.03
450	ADG, kg	1.60	1.66	1.72	1.77	1.83	< 0.01	0.50
451	G:F	0.155	0.156	0.158	0.160	0.162	< 0.01	0.45
452	Feeding value, % <sup>e</sup>		112	112	112	112		

453

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454 <sup>a</sup> WDGS = wet distillers' grains with solubles, MDGS = modified wet distillers' grains with solubles, DDGS = dried distillers' grains  
455 with solubles, DMI = dry matter intake, ADG = average daily gain, and G:F = feed efficiency.

456 <sup>b</sup> Dietary treatment levels (DM basis) of distillers' grains plus solubles (DGS), 0DGS = 0 g/kg DGS, 10DGS = 100 g/kg DGS,  
457 20DGS = 200 g/kg DGS, 30DGS = 300 g/kg DGS, 40DGS = 400 g/kg DGS.

458 <sup>c</sup> Estimation equation linear and quadratic term t-statistic for variable of interest response to DGS level.

459 <sup>d</sup>WDGS data presented are summarized from Bremer et al., 2010.

460 <sup>e</sup> Percent of corn feeding value, calculated from DGS inclusion level feed efficiency relative to 0WDGS feed efficiency, divided by  
461 DGS inclusion.

462 <sup>f</sup> MDGS and DDGS steer performance was scaled to the WDGS intercept for equal comparison across byproduct types. This process  
463 was validated by Nuttelman et al. (2010).

464

465

466 Table 2.

467 Percent reduction in greenhouse gas (GHG) emissions for an equivalent quantity of energy from ethanol relative to gasoline when  
468 accounting for distillers' grains (DGS) moisture content, dietary inclusion level, and livestock type fed.

469 Livestock Type	Beef				Dairy	Swine
470 DGS, g/kg of diet DM <sup>a</sup>	100	200	300	400	100-300	90-270
471 WDGS, GHG % reduction to gasoline <sup>b</sup>	62.4	60.6	58.4	56.7	52.6	---
472 MDGS, GHG % reduction to gasoline <sup>b</sup>	53.9	52.6	50.9	49.7	47.9	---
473 DDGS, GHG % reduction to gasoline <sup>b</sup>	46.1	45.4	44.4	43.9	42.8	42.3

474 <sup>a</sup> DM = dry matter, WDGS = wet distillers' grains with solubles, MDGS = modified wet distillers' grains with solubles, and DDGS =  
475 dried distillers' grains with solubles.

476 <sup>b</sup> Gasoline reference point is 97.7 g CO<sub>2</sub>e/MJ (Liska and Perrin, 2009).

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